

# Investigating the Optimality of Proactive Management of an Invasive Forest Pest

**Craig A. Bond**, Department of Agricultural and Resource Economics, Colorado State University, Fort Collins, CO; **Patricia Champ**, USDA Forest Service, Rocky Mountain Research Station, Fort Collins, CO; **James Meldrum**, Environmental Studies, University of Colorado, Boulder, CO; and **Anna Schoettle**, USDA Forest Service, Rocky Mountain Research Station, Fort Collins, CO

**Abstract**—This paper offers a preliminary investigation into the conditions under which it might be optimal to engage in proactive management of a non-timber forest resource in the presence of an invasive species whose spread is unaffected by management action. Proactive management is defined as treating an uninfected area to encourage healthy ecosystem function, given that the arrival of the invasive is inevitable. Inspired by the problem of white pine blister rust in the Rocky Mountain west of the United States, the model was solved under varying assumptions concerning the scale of management action, benefit and costs, discount rate, and uncertainty of spread. Results showed that proactive strategies tended to be optimal when, all else equal, a) more resources are available for treatment; b) the costs of treatment are rapidly increasing in forest health, or conversely, the benefits of healthy and unhealthy stands are relatively similar; and c) the discount rate is low. The introduction of uncertainty did not significantly affect the likelihood of a proactive management strategy being optimal, but it did show that the conditional probabilities of infection play important role in the decision of which uninfected stand should be treated if a choice is available to the manager.

## Introduction

The emergence of a global economy, associated in large part with increased movement of goods and services, has also increased the probability of non-marketable organisms establishing themselves in areas outside of their native habitat (Mack and others 2000, Mack and Lonsdale 2001). In some cases, economic damages associated with such movement and establishment will be minimal.<sup>1</sup> In others, however, conditions such as a lack of natural enemies for the non-native species and/or a lack of resistance in native organisms to the new species may be sufficient to render significant damages, and earn the label of invasive pest (Schoettle and Sniezko 2007).

Forests are among the ecosystems being impacted by non-native pests and pathogens. Numerous pathogens, non-native arthropod pests and non-native plant species have already disrupted many forest ecosystems throughout North America. Examples include *Cryphonectria parasitica* (Murrill) Barr, the fungal pathogen responsible for chestnut blight of American chestnut trees; *Ophiostoma novo-ulmi* Brasier, the fungal pathogen responsible for the Dutch elm disease of American elm and other native elm species; and *Cronartium ribicola* J.C. Fisch., the fungal pathogen that causes white pine blister rust (WPBR) and cycles between native 5-needle white pines, currants, and gooseberries. Non-native pathogens have

severely reduced some forest species populations, altered forest composition, and threatened the habitats of endangered animals (Liebold and others 1995).

Most invasive species management strategies focus on (1) prevention, (2) early detection and eradication, (3) containment and control, and when those efforts are unsuccessful, (4) mitigation of impacts and (5) restoration of the degraded forest (Schoettle and Sniezko 2007). However, in some cases (such as with WPBR), (1)-(3) have proven challenging, with no effective strategies identified. As such, there is a growing interest in preemptively managing ecosystems to mitigate the potential negative impacts of invasives before significant damage occurs, but without preventing the spread of the pest (usually due to technological or cost reasons). However, only recently have the physical outcomes of these forest management techniques been explored, and the economic conditions under which such “proactive management” is optimal have not been analyzed (Schoettle and Sniezko 2007).

This paper provides a preliminary model that can be used to analyze the conditions under which it might be optimal to pursue a proactive, as opposed to reactive, management strategy in the case of an invasive forest pathogen whose spread cannot be contained. A spatially-explicit stochastic dynamic programming model is developed that tracks the state of each of  $N$  number of stands of a host tree species potentially infected by a damaging invasive species. Subject to the expected evolution of the forest, a manager is assumed to allocate (finite) resources to treat the forest, and can treat any stand in either a proactive (prior to arrival of the invasive) or reactive (after invasive establishment) manner. Results highlight the circumstances under which proactive management is favored, including the physical structure of the forest, stand/forest benefits, management costs, and the probabilities of pathogen spread.

We contribute to the literature in the following ways. First, to our knowledge, there are no published articles in the economics or forestry literature that utilize a dynamic programming methodology to evaluate forest management strategies in the presence of an invasive species. There are, however, a few examples of using these techniques for timber management, including Spring and Kennedy (2005), who examined optimal harvest on multiple stands in the presence of stochastic fire risk and an endangered species in Australia, and Moore and Conroy (2006), who examined silviculture practices for management of old growth forests for habitat purposes in a wildlife refuge in Georgia. Second,

there is little in the economics literature regarding proactive management, perhaps because these strategies are contrary to current conservation approaches that would advocate preservation of so-called “natural” systems, and thus hostile to human intervention into reasonably undisturbed systems. However, proactive management may enable naturalization of the non-native organism while sustaining host populations and ecosystem function (Kilpatrick 2006). Finally, this study contributes to the literature on spatial process in the environmental and resource literature through the incorporation of an explicit spatial structure in the representation of the forest through which an invasive organism moves. In the presence of budget constraints, decisions regarding which stands to manage (either proactively or reactively) will inevitably involve tradeoffs over space as well as time.

## Rationale of Proactive Management: The Case of White Pine Blister Rust (WPBR)

*Cronartium ribicola*, the fungus that causes WPBR, is among the invasive species introductions into North America where containment and eradication efforts have failed (Maloy 1997). It was introduced on the northwest coast of North America from Europe in the early twentieth century, and has since caused a variety of damage to the various species (some keystone) of noncommercial five-needle pines in high elevation North American ecosystems, including foxtail, limber, Rocky Mountain bristlecone, southwestern white, and white-bark pines. WPBR is a lethal disease that causes tree mortality at all life stages, disrupting the regeneration cycle with potentially severe effects on white pine forests.<sup>2</sup> Damages as a result of WPBR infection and tree mortality include effects on various ecosystem components and services such as animal populations (such as Clark’s nutcracker birds, grizzly bears, and red squirrels), watershed production through snow capture, biodiversity and degradation of high-quality recreation opportunities (Petit 2007; Samman and others 2003; Tomback and Kendell 2001; Tomback and others 1995; Mattson 1992; McKinney 2004; Kendell and Arno 1990; McDonald and Hoff 2001). In fact, forests of these types are among the most visited in the country, including those found in the Western region of the National Park system (e.g., Glacier, Yellowstone, and Rocky Mountain National Parks).

The nature of five-needle pine forests suggests that natural evolution of resistance to WPBR is unlikely without intervention<sup>3</sup>, though some natural genetic resistance has been identified in some stands (Sniezko and others 2008; Schoettle and others, Preliminary Overview of the First Extensive Rust Resistance Screening Tests of *Pinus flexilis* and *Pinus aristata*, this proceedings). As such, breeding programs may help to preserve naturally resistant seed stock in high-elevation species, as is being done for commercial species of white pines (McDonald and others 2004). The potential may soon exist for proactive management in which genetically-resistant trees are either directly planted or indirectly encouraged through alternative management actions (stimulating natural regeneration of resistant trees) *prior to* infection (Schoettle 2004a, 2004b;

Schoettle and Sniezko 2007). The rationale behind proactive management, then, is essentially one of “preparing the battlefield” for a transition from an uninfected to infected state. Acting prior to invasion would presumably increase the proportion of genetically resistant trees, thereby reducing impacts on various ecosystem services due to mortality, increasing the probability of a healthy, regenerative system in the long run, and reducing or eliminate the need for reactive management post-invasion. Of course, such management might also be not only directly costly (through management expenditures), but also generate costs (to, say, recreationalists or naturalists) from the disturbance of a previously undisturbed forest. We term such costs “management externalities”.

To date, there has been little information provided to potential forest managers regarding the circumstances under which proactive management might be preferred to the more common reactive strategies (Burns and others 2008). In the following sections, we provide a preliminary model that helps to shed light on these issues. Future research will refine the model using data on non-market benefits of high-elevation forests and the epidemiology of WPBR in the Rocky Mountain region.

## The Dynamic Management Model

### General Description

We assume that a resource manager has responsibility over a forest threatened by a non-native species whose spread cannot be arrested through any management action (a circumstance such as WPBR). As in Spring and Kennedy (2005), the forest is composed of  $N$  stands, with the state of each stand in time period  $t$  represented by one of a countable number of states representing a) the health of the stand (or level of ecosystem services provided by the stand) and b) the status of the stand as “treated” or “untreated”. An untreated stand, once infected by the invasive pest and left untreated, will dynamically evolve such that mortality increases (ecosystem services decrease) until a terminal level is reached and maintained throughout the infinite time horizon of the problem. Once treated, perhaps by either planting resistant seedlings or otherwise encouraging reproduction of resistant biomass, a stand recovers until it reaches a relatively healthy terminal state, where it remains for the remainder of the problem. Treatment thus does not eliminate invasive spread, but minimizes long-run impact through the addition of resistant trees in the spirit of Kilpatrick (2006). Treatment of an uninfected stand results in a transition to a healthy state with probability one, in accordance with the rest of the forest dynamics detailed below.

Managers may treat any stand at any time, but are subject to a budget constraint that limits the number of stands treated in any one decision period. For simplicity, we assume only one treatment alternative whose success is certain (though this is fairly easily relaxed), and per-stand treatment costs are assumed to decrease with tree mortality (increase with ecosystem service provision). As noted above, spread of the invasive species is assumed not to depend on management actions, and is directional and potentially probabilistic in its spread (as

in the case of WPBR). Ecosystem service benefits from the physical state of each stand are assumed to be homogeneous and decreasing in stand mortality, and total net benefits from the forest are additive across stands. The manager is assumed to maximize the net present value of the expected net benefits from stand treatment over an infinite time horizon, subject to the spread and damage caused by the invasive species and the budget constraint.

### Forest Dynamics

The model of the forest is cellular and spatial in nature, with  $N=4$  stands. At any time  $t$ , each stand  $X_i$ ,  $i = 1, \dots, N$ , is assumed to be in one of  $S = 7$  discrete states representing the overall health of the stand and the treatment status of it. Overall, there are three health states corresponding to ecosystem service provision (healthy, moderately healthy, and not healthy) and two treatment states (treated and untreated) for stands that have been infected by the invasive, plus one more state representing a healthy stand that has not yet been exposed to the non-native pathogen. The total number of potential states of the forest is thus  $S^N = 7^4 = 2,401$ , which illustrates the necessity of restricting attention to four stands using standard discrete-space numeric dynamic programming techniques.<sup>4</sup>

The states of each stand are defined categorically, where  $X_i = 0$  implies lack of invasive establishment on an untreated stand. Let  $\tau_i$  be an indicator variable that signifies if stand  $i$  has ever been treated, and restrict attention to stands where the invasive has been established. As such, untreated stands can take on states

$$X_i = \begin{cases} 1 & \text{if } \tau_i = 0 \text{ and stand } i \text{ is healthy} \\ 2 & \text{if } \tau_i = 0 \text{ and stand } i \text{ is moderately healthy} \\ 3 & \text{if } \tau_i = 0 \text{ and stand } i \text{ is not healthy} \end{cases} \quad (1)$$

Once treatment has occurred, the three potential states are

$$X_i = \begin{cases} 4 & \text{if } \tau_i = 1 \text{ and stand } i \text{ is healthy} \\ 5 & \text{if } \tau_i = 1 \text{ and stand } i \text{ is moderately healthy} \\ 6 & \text{if } \tau_i = 1 \text{ and stand } i \text{ is not healthy} \end{cases} \quad (2)$$

Note that by assumption, an uninfected or “just infected” stand (states 0 and 1) immediately transitions to the terminal healthy state (state 4) if treated.

State transitions in time  $t+1$  depend on the initial state of the stand at time  $t$  (namely  $x_{it}$ ), the value of the treatment control variable for that stand ( $u_{it} = 1$  if treated), and in the case of an uninfected stand, the event of stand infection and establishment, denoted by the event indicator  $\phi_i = 1$ . The state transitions are thus defined as

$$x_{it+1}(x_{it}, \phi_{it}, u_{it}) = \begin{cases} 0 & \text{if } x_{it} = 0 \text{ and } \phi_{it} = 0 \text{ and } u_{it} = 0 \\ 1 & \text{if } x_{it} = 0 \text{ and } \phi_{it} = 1 \text{ and } u_{it} = 0 \\ 4 & \text{if } x_{it} = 0 \text{ and } \phi_{it} = 0 \text{ and } u_{it} = 1 \\ x_{it} + 1 & \text{if } 0 < x_{it} < 3 \text{ and } u_{it} = 0 \\ x_{it} + 4 & \text{if } 0 < x_{it} < 3 \text{ and } u_{it} = 1 \\ x_{it} - 1 & \text{if } 4 < x_{it} \leq 6 \\ 3 & \text{if } x_{it} = 3 \text{ and } u_{it} = 0 \\ 6 & \text{if } x_{it} = 3 \text{ and } u_{it} = 1 \\ 4 & \text{if } x_{it} = 4 \end{cases} \quad (3)$$

Note that state 3 (unhealthy stand) is a terminal state for untreated regions, while state 4 (healthy stand) is a terminal state for treated regions. Assuming that the effects of treatment are certain and there are no other exogenous threats to the forest (for example, fire, climate change ...), the only stochastic element in the model is the infection and establishment event  $\phi_{it} = 1$ . We turn to considerations of this variable in the next subsection.

### Probabilities of Stand Infection and Spatial Forest Structure

The spatial configuration of the forest is represented by a  $N \times N$  matrix  $\mathbf{z}$  with elements  $z_{ij} = (0,1)$ . For row  $i$ , a non-zero element in position  $j$  indicates that an infected neighbor  $j$  increases the probability of infection of stand  $i$  in the following period. Similarly, for column  $j$ , a non-zero element in row  $i$  indicates that stand  $i$  is more at risk once  $j$  is infected. As such, through specification of this matrix, a “directionality” of spread can be modeled. For example, suppose that spread is deterministic in a southeast direction (including due east and due south), in the sense that once a neighbor to the north or west of stand  $i$  is infected in time  $t$ , then stand  $i$  will become infected in time  $t+1$  with a probability of one, and otherwise will not be infected. Further assume that are stands arranged in a rectangular formulation such that stand 1 is to the northwest, stand 2 is northeast, stand 3 is in the southwest, and stand 4 is in the southeast. The matrix  $\mathbf{z}$  is thus defined as

$$\mathbf{z} = \begin{bmatrix} 0 & 0 & 0 & 0 \\ 1 & 0 & 0 & 0 \\ 1 & 0 & 0 & 0 \\ 1 & 1 & 1 & 0 \end{bmatrix}, \quad (4)$$

so that, for example, stand 4 will be infected in  $t+1$  if any of stands 1, 2, or 3 are infected in time  $t$  (row 4), but the infection status of stand 2 only affects the probabilities associated with stand 4 (2<sup>nd</sup> column).

In general, we assume that the probabilities associated with establishment of the invasive on a given stand are a function of the number of infected neighboring stands as defined by the matrix  $\mathbf{z}$ . Let  $s_{ij} = 1$  if  $x_j > 0$ , 0 otherwise, and define the number of infected neighboring stands for stand  $i$  as  $\bar{n}_i = \sum z_{ij} \cdot s_{ij}$ , with  $0 \leq \bar{n}_i \leq 3$ . The infection and establishment event, then, is a function of the spatial structure of the



**Table 1.** Stand infection probabilities as a function of number of infected neighbors, deterministic and stochastic cases.

# of infected neighboring stands ( $\bar{n}_i$ )	$\Pr(\emptyset_i   \bar{n}_i(x, z))$	
	Deterministic	Stochastic
0	0.0	0.1
1	1.0	0.6
2	1.0	0.8
3	1.0	0.9

forest and the states of the surrounding stand, and the associated probabilities, namely  $\Pr(\emptyset_i | \bar{n}_i(x, z))$  are given in Table 1. Note that for this paper, the probabilities are illustrative, and not empirically based.

Using these probabilities, we define  $\Pr(x_{ij}^+ | x_i, \bar{n}_i(x, z), u_i)$  to be the probability of a stand transitioning from state  $x_i$  to state  $x_{ij}^+$  conditional on the state of the forest and the control chosen. Of the  $S^N$  potential states in the model, then, the transitions associated with  $(S-1)^N$  are deterministic. In the case presented here, this is approximately 54 percent of all possible starting states.

### Economic Parameters

Table 2 reports information about the benefits and costs associated with forest management. We assume that in each (multi-year) period, benefits from the forest are the net present value of the sum of stand-level ecosystem service benefits, which are increasing with the health of each stand. We denote these as  $f(x_i)$ . Treatment costs  $c(u_i, x_i)$  are incurred only in the current period, and are decreasing with the health of each stand due to ease of management and the potential for management externalities.

The manager is assumed to be constrained in action due to budget, and as such can only treat a limited number of stands per period.<sup>5</sup> As such, the control set  $U$  is defined directly from this constraint. For example, if the budget is one stand per year, then the number of elements in  $U$  is five, corresponding to treating each individual stand plus not treating any. If, however, two stands may be treated in the same time period, then the control set is augmented to include eleven possible stand combinations.

**Table 2.** Net present value of benefits and costs for forest stand states per time period, baseline scenario.

State of stand $x_i$	Description	Per-stand benefits $f(x_i)$	Per-stand treatment costs $c(u_i, x_i)$
0	Uninfected and not established	10	7
	Uninfected, healthy		
1	Infected and established	10	7
	Infected and healthy		
2	Infected and moderately healthy	5	5
3	Infected and not healthy	0	2
4	Treated and healthy	10	7
5	Treated and moderately healthy	5	5
6	Treated and not healthy	0	2

Collecting these assumptions and placing them in the framework of a dynamic programming problem, the discrete-time Bellman equation characterizing the problem is

$$\begin{aligned}
 V(\mathbf{x}) &= \max_{\mathbf{u} \in U} \{ \sum_i [f(x_i) - c(u_i, x_i)] + \beta E[V(\mathbf{x}^+(\mathbf{x}, \boldsymbol{\varphi}, \mathbf{u}))] \} \\
 &= \max_{\mathbf{u} \in U} \{ \sum_i [f(x_i) - c(u_i, x_i)] \\
 &\quad + \beta \sum_{j=1}^{S^N} [\Pr(\mathbf{x}_j^+ | \mathbf{x}, \bar{\mathbf{n}}(\mathbf{x}, \mathbf{z}), \mathbf{u}) V(\mathbf{x}_j^+(\mathbf{x}, \boldsymbol{\varphi}, \mathbf{u}))] \}, \tag{5}
 \end{aligned}$$

where  $\mathbf{x}^+(\mathbf{x}, \boldsymbol{\varphi}, \mathbf{u})$  is the vector of state transition equations defined in (3),  $\Pr(\mathbf{x}_j^+ | \mathbf{x}, \bar{\mathbf{n}}(\mathbf{x}, \mathbf{z}), \mathbf{u})$  is the probability of transition from state  $\mathbf{x}$  to  $\mathbf{x}_j^+$ , defined as the product of the stand level probabilities  $\Pr(x_{ij}^+ | x_i, \bar{n}_i(x, z), u_i)$ , and  $\beta$  is the discount factor, suitably defined to reflect the number of years assumed between each time period.

The model was coded and solved numerically in MATLAB using the default policy iteration method of the CompEcon toolbox in Miranda and Fackler (2002). Given a particular parameterization of the model (including the probabilities, benefits and costs of each stand in each state, and the discount factor), the solution to (5) allows for recovery of the optimal management plan that maximizes the net present value of the entire forest (four stands) over an infinite time horizon using standard dynamic programming techniques (see Miranda and Fackler 2002). These optimal strategies are functions of the states of the system (defined as the health of all four stands), and take the form of a four by one vector that indicates treatment or non-treatment of each stand in each state. For the purposes of this study, treatment of a stand before infection is termed proactive.

## Results

### Optimal Deterministic Policies

Optimal policies for a sample of starting states under two budget constraints (a maximum of one stand treated per decision period and a maximum of two stands treated per decision period) are presented in Table 3, assuming deterministic invasive species spread in the southeast direction with stands one and two to the north and stands three and four to the south arranged in a rectangular fashion (see Figure 1). The discount factor is assumed to be 0.9. Note that “do nothing” is an admissible management strategy in all cases; as such, the optimal results reported below are superior to this option.

Under the baseline parameterization, and considering the case of a maximum of one treated stand per period, there are 1,105 forest configurations in which proactive management, defined relatively strictly as treating an uninfected, previously untreated stand, is feasible.<sup>6</sup> Of this set, approximately 13 percent (145) of the optimal management strategies could be classified as proactive. The large majority of these occur when the infection threat is immediate (i.e., a stand to the northwest of an uninfected stand is infected), and the other

**Table 3.** Optimal policies for selected starting states and budget constraints, deterministic model.

Starting States				Optimal Treated Stands and Proactive Indicator			
				max 1 treated		max 2 treated	
Stand 1	Stand 2	Stand 3	Stand 4	Treated Stand	Proactive?	Treated Stands	Proactive?
0	0	0	0	none	no	none	no
1	0	0	0	1	no	1,2	yes
1	1	0	0	1	no	1,2	no
1	4	0	0	1	no	1,3	yes
2	0	0	0	2	yes	1,2	yes
2	4	4	1	3	n/a	3,4	n/a
5	4	4	1	4	n/a	4	n/a
6	4	4	5	none	n/a	none	n/a
4	4	4	4	none	n/a	none	n/a

stands are either uninfected or have already been treated, and thus are in states 0 or 4 through 6. Intuitively, this makes sense as the opportunity costs of treating a stand proactively in this case are small, given that the remainder of the forest is relatively protected and increasing in health. If, however, at least one stand is actively degrading or degraded (states 1-3), it is generally optimal to treat one of these stands in a reactive fashion (though the specifics depend on the relative states of each degrading stand and the potential for damage through spread). One exception to this prescription is if exactly one of the stands is only moderately healthy (state = 2) and the only other infected stand has been treated. In this case, the optimal strategy is to proactively treat the northeast-most uninfected stand. Presumably, this result occurs as a result of the interaction between the opportunity costs of treatment and the fact that treatment costs for the moderately infected stand will fall enough such that it pays to wait to treat. We further explain the incentives below.

If the budget constraint is relaxed to accommodate treatment of up to two stands per time period, then the percentage of times it is optimal to pursue proactive strategies increases to 41%, more than three times the one-stand per time period number. This set of proactive strategies generally includes cases where if there are two or more stands infected, at least one has already been treated. Given the flexibility inherent in this parameterization of the problem, the spatial dimension is more apparent as well. For example, a manager will generally treat degrading cells to the northwest, *ceteris paribus*, through s/he still must trade off the potential for spread and increased future damage with the cost decrease (and own-stand damage increase) if treatment does not occur. As such, we conclude that proactive management under this deterministic directional spread scenario is generally favored as resource constraints are relaxed, but not at the expense of reactive management when multiple stands are degrading. However, this is but one set of benefit and cost schedules, suggesting an analysis of the effects of these measures at the margin is appropriate.

### Effects of Benefits and Costs

Of course, the tradeoffs involved in dynamic forest management in the presence of an invasive species are in large part

determined by the marginal benefits and costs of treatment, which in turn depend on both spatial and temporal features. We now turn to the effects of shifting the relative benefit and cost schedules associated with forest stands in order to determine their effects, and thus provide some sensitivity analysis of the results.

To illustrate, we run an experiment which doubles the cost of treatment in healthy stands and cuts the cost of treatment in unhealthy stands by half (the “high cost” scenario), while keeping costs for the moderately healthy stands the same in the two-stand constrained deterministic spread model. Thus, we have increased the marginal costs of treating a healthy forest, perhaps mirroring a case of relatively severe management externalities.

Following our earlier analysis, proactive strategies are now optimal for almost 57% (626/1105) of possible cases, despite the increase in treatment costs for uninfected and healthy stands. Part of the reason can be seen in from the difference in strategies from case (a) when  $x_a = [1 \ 0 \ 0 \ 0]'$  and case (b) where  $x_b = [1 \ 1 \ 0 \ 0]'$ , as seen in Table 4. When the cost of treatment for healthy stands is relatively low, initial optimal treatment  $u_a^{low} = [treat \ 1 \& 2]$ , but when it is relatively high, initial optimal treatment changes to  $u_a^{high} = [treat \ 2 \& 3]$ . Similarly, for  $x_b$ ,  $u_a^{low} = [treat \ 1 \& 2]$  and  $u_a^{high} = [treat \ 3 \& 4]$ . Note that in case *a*, both scenarios involve proactive management, while in case *b*, only  $u_a^{high}$  treats (both) uninfected stands.

This result cannot simply be explained by a change in the relative costs across cells, as treatment costs are homogeneous across all four stands. As such, the answer must lie with the opportunity costs of treatment. Advancing the system in case (a) according to the optimal policy,  $x_a^{+low} = [5 \ 4 \ 1 \ 1]'$  and  $x_a^{+high} = [2 \ 4 \ 4 \ 1]'$ , with corresponding policies at these new states defined by  $u_a^{+low} = [treat \ 3 \& 4]$  and  $u_a^{+high} = [treat \ 1 \& 4]$ . Following the paths to their terminal states of  $x^\infty = [4 \ 4 \ 4 \ 4]'$ , as in Table 4, it is clear that the *low* takes three decision periods to reach  $x^\infty$ , while the *high* case takes four. The reason is that in the *high* case, the marginal benefit from the treatment cost reduction outweighs the discounted marginal reduction in benefits from allowing stand 1 to devolve into an unhealthy state, and then recovering once treated. Thus, the manager prefers what we might call a “purely” proactive strategy in

**Table 4.** Sample simulations under alternative treatment cost assumptions, deterministic, two-stand constraint model.

Time Period	Low Cost Scenario				Case <i>a</i> High Cost Scenario				High Cost Scenario Using Low-Cost Policy			
	Forest State	Treated Stands	Benefits - Costs	NPV	Forest State	Treated Stands	Benefits - Costs	NPV	Forest State	Treated Stands	Benefits - Costs	NPV
0	[1 0 0 0]	1,2	26	26.00	[1 0 0 0]	2,3	12	12.00	[1 0 0 0]	1,2	12	12.00
1	[5 4 1 1]	3,4	21	18.90	[2 4 4 1]	1,4	16	14.40	[5 4 1 1]	3,4	7	6.30
2	[4 4 5 5]	n/a	30	24.30	[6 4 4 5]	n/a	25	20.25	[4 4 5 5]	n/a	30	24.30
3	[4 4 4 4]	n/a	40	29.16	[5 4 4 4]	n/a	35	25.52	[4 4 4 4]	n/a	40	29.16
4	[4 4 4 4]	n/a	40	<u>26.24</u>	[4 4 4 4]	n/a	40	<u>26.24</u>	[4 4 4 4]	n/a	40	<u>26.24</u>
		Total		124.60		Total		98.41		Total		98.00

  

Time Period	Low Cost Scenario				Case <i>b</i> High Cost Scenario				High Cost Scenario Using Low-Cost Policy			
	Forest State	Treated Stands	Benefits - Costs	NPV	Forest State	Treated Stands	Benefits - Costs	NPV	Forest State	Treated Stands	Benefits - Costs	NPV
0	[1 1 0 0]	1,2	26	26.00	[1 1 0 0]	3,4	12	12.00	[1 1 0 0]	1,2	12	12.00
1	[5 5 1 1]	3,4	16	14.40	[2 2 4 4]	1,2	20	18.00	[5 5 1 1]	3,4	2	1.80
2	[4 4 5 5]	n/a	30	24.30	[6 6 4 4]	n/a	20	16.20	[4 4 5 5]	n/a	30	24.30
3	[4 4 4 4]	n/a	40	29.16	[5 5 4 4]	n/a	30	21.87	[4 4 4 4]	n/a	40	29.16
4	[4 4 4 4]	n/a	40	<u>26.24</u>	[4 4 4 4]	n/a	40	<u>26.24</u>	[4 4 4 4]	n/a	40	<u>26.24</u>
		Total		120.10		Total		94.31		Total		93.50

Low cost scenario: Treatment costs = \$7 for healthy, \$5 for moderately healthy, \$2 for unhealthy

High cost scenario: Treatment costs = \$14 for healthy, \$5 for moderately healthy, \$1 for unhealthy

Discount factor = 0.90

period one, but does so, perhaps counter intuitively, to capture the “benefits” of stand degradation.

Turning to case *b*, we see a very similar result, as the manager prefers to engage in a proactive strategy to protect stands 3 and 4 in the first period, while allowing for stands 1 and 2 to degrade in order to take advantage of the relative cost savings offered by treating partially healthy forests. These savings dominate the decision despite the additional expense of losing benefits in period two (after the second control decision), relative to the *low* case, as a result of two unhealthy treated stands that take an extra period to return to health.

We have thus illustrated that proactive strategies tend to be favored when the costs of stand treatment are increasing relatively rapidly in stand health, and conversely, then, when the benefits of stand health are relatively unresponsive to degradation. Put another way, the greater the change in net benefits as forest health changes, the more likely is proactive management to be optimal in a dynamic spatial setting, as the presence of “substitute” stands allows managers to take advantage of the cost savings resultant from degradation. Given the role that future damages play in the analysis, however, we now turn to the effect of the discount rate on the solution to the problem.

### Effect of Discount Rate

The discount factor provides a relative weighting between the (unspecified) time period between which decisions regarding treatment are made and the forest stands evolve. The discount factor is defined as  $\beta = \frac{1}{(1+r)}$  where  $r$  is the discount rate that represents the opportunity cost of capital, or conversely, the rate at which the next best alternative asset appreciates. In economic theory, the discount rate is used to represent the idea that one dollar of benefits today is preferred to one dollar of benefits in a future time period, as there is an intertemporal opportunity cost to waiting.

The baseline analysis assumed a discount factor of  $\beta=0.9$ . Without greater biological detail, it is hard to determine if such a weighting is appropriate for all scenarios. On the one hand, the length of time it takes species such as five-needle pines to grow and evolve might suggest that the discount factor should be lower; on the other hand, intergenerational equity and other concerns provide an argument that the discount factor should be relatively close to one (Spring and Kennedy 2005; Weitzman 2001). In order to investigate the effects of the discount rate, additional scenarios were analyzed as the discount factor decreased (less weight on the future). One would suspect that as the present was favored, the incentives for proactive management would decrease as the marginal benefits of treating an individual stand would decrease. In fact, this is exactly the case, and in some cases, is quite dramatic. For example, if the discount factor is 0.5 under the two-stand constraint, then the optimal strategy is to treat only completely degraded stands once that state is reached, and do nothing to any other stand in any other state. As such, the percentage of potential proactive management occasions that are optimal is zero. At  $\beta=0.65$ , this percentage increases to a very small one half of one percent (all cases where stand 1, which is positioned to spread the invasive to all other stands, is infected), and when  $\beta=0.70$  and higher, the result is identical to the baseline scenario.

As such, so long as the discount rate (factor) is sufficiently low (high), proactive management strategies are part of the optimal forest management plan. In the cases considered here, there is a fairly narrow range with  $.60 < \beta < .70$  over which the optimal policies are affected, and tend to favor proactive strategies only when the spread potential for the invasive species is high and the forest is generally healthy. This corresponds to a situation in which a low weight placed on future outcomes is outweighed by the damage caused from increased invasive spread.

## Effects of Uncertainty

In addition to the deterministic scenarios analyzed above, the model was also solved taking into account a probabilistic establishment regime for the invasive (see Table 1), but maintaining all other baseline scenario parameters for the two-stand constrained problem. In general, this scenario assumes that the threat of the invasive to an uninfected stand is increasing in the number of infected stands that have the ability to threaten it (in the sense of the matrix  $\mathbf{z}$ ). In addition, there is an external threat in that the forest in the state  $[0\ 0\ 0\ 0]$  can become infected (in this case, with a probability of .4). For simplicity, the manager is assumed to maximize the expected net present value of profits, and thus is risk neutral in preferences.

Results of this exercise reveal that only small changes in optimal policy rules occur as a result of the uncertainty over spread.<sup>7</sup> In each case, it involves two infected stands with one treated, but the other two are undisturbed and must include stand 4 (the most threatened due to the directional nature of the spread). As direct result of the differential in probabilities of potential spread between the two stands, it is always optimal in the stochastic case to treat the “more threatened” stand 4, primarily as a direct result of the differential in probabilities of potential spread between the two stands. In the deterministic case, given the  $\mathbf{z}$  matrix, the manager is indifferent between which stands to treat, as the probabilities related to spread are identical. As a result, there is no effect in the frequency of optimal proactive management over the deterministic case; rather, this result serves to guide the choice of stands to proactively manage, if there is indeed such a choice.

## Discussion and Conclusions

This paper offers a preliminary investigation into the conditions under which it might be optimal to engage in proactive management of a non-timber forest resource in the presence of an invasive species whose spread is unaffected by management action. Although contrary to current practice, proactive management is defined as treating an uninfected area to encourage healthy ecosystem function, given that the arrival of the invasive is inevitable. The model is inspired by the problem of white pine blister rust (WPBR) in Whitebark Pine in the Rocky Mountain west of the United States, which has severely impacted Glacier National Park, and is currently threatening Yellowstone and Rocky Mountain National Park, among other public lands.

The model was solved under varying assumptions concerning the potential scale of management action (through the budget constraint), the benefit and cost schedules associated with the forest resource, the discount rate, and the level of uncertainty of spread. The primary management implications are that, all else equal, proactive management strategies are preferred when: a) more resources are available for treatment (a greater number of stands can be treated in any one decision period); b) the costs of treatment are rapidly increasing in forest health, or conversely, the benefits of healthy and unhealthy stands are relatively similar; and c) the discount factor (rate) is high (low), implying a relatively high weight on the future.

Additionally, although the introduction of uncertainty did not significantly affect the likelihood of a proactive management strategy being optimal, it did show that the conditional probabilities of infection play important role in the decision of which uninfected stand should be treated if a choice is available to the manager. At a more basic level, the results of the exercise can aid in developing optimal management plans so long as the model can be parameterized.

Although relatively simple, the model presented here should help managers understand the incentives related to non-timber forest management in the presence of an unavoidable and unalterable threat from an invasive species. Given a parameterization based in empirical data, this framework can be used to define optimal management plans given the state of a particular set of stands, and when and where proactive (and, indeed, reactive) management is preferred. Furthermore, it could also be used to evaluate the differences in discounted net benefits between treatment plans, though this is not explored in the current paper.

That said, future research can do much to clarify and augment the conclusions reported here. For example, improved parameterizations for a given circumstance, including the economic and biological/epidemiological representations of the system based on collected data, could assuage concerns about arbitrary assumptions. This includes not only state-space representation of the forest, but the number of potential management units as well. Similarly, managers have multiple treatment strategies available (planting, burning, both...), with outcomes of any strategy likely uncertain, with potentially varying streams of benefits and costs over time. As the modeling effort becomes more complex and thus more reflective of the system it represents, the results presented here can be used to verify and validate future results, help managers draw conclusions about the general conditions under which proactive and reactive management strategies are optimal, and inform about other similar processes and problems, such as the spread of infectious disease.

## References

- Bertsekas, D. P.; Tsitsiklis, J. N. 1996. *Neuro-Dynamic Programming*. Belmont, MA: Athena Scientific.
- Brustein, F. C.; Yamaguchi, D. K. 1992. The oldest known Rocky Mountain bristlecone pines (Engelm.) Arctic and Alpine Research. 24(3): 253-256.
- Burns, K. S., A. W. Schoettle, W. R. Jacobi, and M. F. Mahalovich. 2010. White pine blister rust in the Rocky Mountain Region and options for management. USDA Forest Service RMRS-GTR 206. 26p. Available at: [http://www.fs.fed.us/rm/pubs/rmrs\\_gtr206.html](http://www.fs.fed.us/rm/pubs/rmrs_gtr206.html).
- Curry, D. R. 1965. An ancient bristlecone pine stand in eastern Nevada. *Ecology*. 46: 564-566.
- Kendall, K. C.; S. F. Arno. 1990. Whitebark pine—an important but endangered wildlife resource. In: Schmidt W. C. and K. J. McDonald (eds). *Proceedings—Symposium on whitebark pine ecosystems: Ecology and management of a high-mountain resource*. General Technical Report INT-270, Ogden, UT: USDA Forest Service, Intermountain Research Station. pp 264-273.
- Kilpatrick, A. M. 2006. Facilitating the evolution of resistance to avian malaria in Hawaiian birds. *Biological Conservation*. 128: 475-485.



- Liebold, A. M., W. L. MacDonald, D. Bergdahl, and V. C. Mastro. 1995. Invasion by exotic forest pests: A threat to forest ecosystems. *Forest Science*. 41(2): 1-49.
- Mack, R. N., D. Simberloff, W. M. Lonsdale, H. Evans, M. Clout, and F. A. Bazzaz. 2000. Biotic invasions: causes, epidemiology, global consequences, and control. *Ecological Applications*. 10: 689-710.
- Mack, R. N.; W. M. Lonsdale. 2001. Humans as global plant dispersers: getting more than we bargained for. *BioScience*. 51: 95-102.
- Maloy, O. C. 1997. White pine blister rust in North America: a case history. *Annual Review of Phytopathology*. 35: 87-109.
- Mattson, D. J. 1992. Yellowstone grizzly bear mortality, human habituation and whitebark pine seed crop. *Journal of Wildlife Management*. 56: 432-442.
- McDonald, G. I.; R. J. Hoff. 2001. Blister rust: an introduced plague. In: Tomback D. F., S. F. Arno, and R. E. Keane, eds. *Whitebark Pine Communities*. Washington, DC: Island Press. Pp. 193-220.
- McDonald, G., P. Zambino, and R. Sniezko. 2004. Breeding rust-resistant five-needle pines in the western United States: lessons from the past and a look to the future. In: Sniezko, R. A., S. Samman, S. E. Schlarbaum, and H. B. Kriebel, eds. *Breeding and genetic resources of five-needle pines: growth, adaptability, and pest resistance*; 2001 July 23-27; Medford, OR, USA.
- Miranda, M. J.; P. L. Fackler. 2002. *Applied Computational Economics and Finance*. Cambridge, MA: MIT Press.
- Party 2.02.15. Proceedings RMRS-P-32. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station. Pp 28-50.
- McKinney, S. T. 2004. Evaluating natural selection as a management strategy for restoring whitebark pine. Master of Arts Thesis, University of Colorado: Denver, CO.
- Moore, C. T.; M. Conroy. 2006. Optimal regeneration planning for old-growth forest: addressing scientific uncertainty in endangered species recovery through adaptive management. *Forest Science*. 52(2): 155-172.
- Petit, C. (Jan 30 2007) In the Rockies, pines die and bears feel it. *New York Times*. Available at: <http://www.nytimes.com/2007/01/30/science/30bear.html?ex=1327813200&en=9ba8339901252668ei=5088partner=rssnytemc=rss#>.
- Samman S., J. W. Schwandt, and J. L. Wilson. 2003. Managing for healthy white pine ecosystems in the United States to reduce the impacts of white pine blister rust. Forest Service Report R1-03-118. Missoula, MT: USDA, Forest Service. 10 p.
- Schauer, A. J., A. W. Schoettle, and R. L. Boyce. 2001. Partial cambial mortality in high-elevation *Pinus aristata* (Pinaceae). *American Journal of Botany*. 88: 646-652.
- Schoettle, A. W. 1994. Influence of tree size on shoot structure and physiology of *Pinus contorta* and *Pinus aristata*. *Tree Physiology*. 14: 1055-1068.
- Schoettle, A. W. 2004a. Ecological roles of five-needle pine in Colorado: Potential consequences of their loss. In: Sniezko, R. A., S. Samman, S. E. Schlarbaum, and H. B. Kriebel, eds. *Breeding and genetic resources of five-needle pines: growth, adaptability, and pest resistance*; 2001 July 23-27; Medford, OR, USA. IUFRO Working Party 2.02.15. Proceedings RMRS-P-32. Fort Collins, CO: USDA Forest Service, Rocky Mountain Research Station. Pp 124-135. Available at: [http://www.fs.fed.us/rm/pubs/rmrs\\_p032/rmrs\\_p032\\_124\\_135.pdf](http://www.fs.fed.us/rm/pubs/rmrs_p032/rmrs_p032_124_135.pdf).
- Schoettle, A. W. 2004b. Developing proactive management options to sustain bristlecone and limber pine ecosystems in the presence of a non-native pathogen. In: Shepperd, W. D., L. G. Eskew, compilers. *Silviculture in special places: Proceedings of the National Silviculture Workshop*; 2003 September 8-11; Granby, CO. Proceedings RMRS-P-34. Fort Collins, CO: USDA Forest Service, Rocky Mountain Research Station. Pp 46-155.
- Available at: [http://www.fs.fed.us/rm/pubs/rmrs\\_p034/rmrs\\_p034\\_146\\_155.pdf](http://www.fs.fed.us/rm/pubs/rmrs_p034/rmrs_p034_146_155.pdf).
- Schoettle, A. W.; R. A. Sniezko. 2007. Proactive intervention to sustain high elevation pine ecosystems threatened by white pine blister rust. *Journal of Forest Research*. 12: 327-336.
- Schulman, E. 1958. Bristlecone pine, oldest known living thing. *National Geographic Magazine*. 113: 355-372.
- Sniezko, R. A.; A. Kegley, R. Danchok, A. W. Schoettle, K. S. Burns, and D. Conklin. 2008. *Cronartium ribicola* resistance in whitebark pine, southwestern white pine, limber pine and Rocky Mountain bristlecone pine—preliminary screening results from first tests at Dorena GRC. In: McWilliams, Michael; Palacios, Patsy, comps. *Proceedings of the 55th Annual Western International Forest Disease Work Conference*; October 15-19, 2007; Sedona, AZ. Logan: Utah State University: 84-86. <http://www.fs.fed.us/foresthealth/technology/wif/proceedings/WIFDWC2007.pdf>
- Spring, D. A.; J. O. S. Kennedy. 2005. Existence value and optimal timber-wildlife management in a flammable multistand forest. *Ecological Economics*. 55: 365-379.
- Tomback, D. F.; K. C. Kendell. 2001. Biodiversity losses: the downward spiral. In: Tomback, D. F., S. F. Arno, and R. E. Keane (eds). *Whitebark Pine Communities*. Washington, DC: Island Press. Pp 243-262.
- Tomback, D. F., J. K. Clary, J. Koehler, R. J. Hoff, and S. F. Arno. 1995. The effects of blister rust on post-fire regeneration of whitebark pine: the Sundance burn of Northern Idaho (USA). *Conservation Biology*. 9: 654-664.
- Weitzman, M. L. 2001. Gamma discounting. *American Economic Review*. 91(1): 260-271.

---

#### ENDNOTES

- <sup>1</sup> Such damages can be to marketable and/or non-marketable ecosystem services.
- <sup>2</sup> Some infected areas in the American West have seen mortality of up to 90%.
- <sup>3</sup> Individuals within these species can live for 1,000-4,500 years, can thrive in harsh environments, and are not frequently disturbed through stochastic events such as fire (Schoettle 1994; Schoettle and Rochelle 2000; Schauer et al. 2001; Schulman 1958; Curry 1965; Brustein and Yamaguchi 1992).
- <sup>4</sup> For larger state spaces, more advanced techniques (rollout strategies, temporal difference learning, etc.) can be used to approximate the optimal solution. See, e.g., Bertsekas and Tsitsiklis (1996).
- <sup>5</sup> Given this assumption, the interpretation of the budget constraint should not be strictly monetary. Rather, one might interpret it as a binding constraint on additional resources, such as labor or capital.
- <sup>6</sup> Given the state transition structure assumed here, it might be logical to term treatment of infected, healthy stands (state 1) as proactive; but we choose not to in order to shed light on primarily “preventative” management options.
- <sup>7</sup> We expect no difference in policy rules where proactive management is not possible, as these transitions are deterministic by assumption.

---

The content of this paper reflects the views of the author(s), who are responsible for the facts and accuracy of the information presented herein.

---